

# **Journal of Biodiversity and Environmental Sciences (JBES)** ISSN: 2220-6663 (Print) 2222-3045 (Online) Vol. 2, No. 12, p. 24-40, 2012 http://www.innspub.net

**RESEARCH PAPER** PEN ACCESS

# species of katerniaghat Recongregation of tree wildlife sanctuary, Uttar Pradesh, India

Omesh Bajpai<sup>1,2</sup>, Anoop Kumar<sup>1</sup>, Ashish K. Mishra<sup>1</sup>, Nayan Sahu<sup>1</sup>, Jitendra Pandey<sup>2</sup>, Soumit K. Behera<sup>1</sup>, Lal Babu Chaudhary<sup>1</sup>

Plant Diversity, Systematics and Herbarium Division, CSIR-National Botanical Research Institute, Rana Pratap Marg, Lucknow, Uttar Pradesh-226 001, India

<sup>2</sup>Centre of Advanced Study in Botany, Banaras Hindu University, Varanasi, Uttar Pradesh-221 005, India

Received: 28 November 2012 Revised: 15 December 2012 Accepted: 16 December 2012

**Keywords:** Tree diversity, recongregation, clustering, PCA, katerniaghat wildlife sanctuary.

#### Abstract

The study highlights the association pattern of tree species in Katerniaghat Wildlife Sanctuary (KWS) a tropical moist deciduous forest of Terai region of Uttar Pradesh, India. In the present study the ecological exploration has been done to dig out the current statuse of the forest types in the sanctuary area. The cluster and PCA analyses grouped the forests of KWS into three major forest types: sal forest, miscellaneous forest and teak plantation. The different forest type exhibit differences in tree dominance and diversity: sal forest is dominated by Shorea robusta, miscellaneous forest by Mallotus philippensis and teak plantation by Tectona grandis. The tree species richness has been found 48.07 species ha<sup>-1</sup> in teak plantation, 52.63 in miscellaneous forest and 55.35 in sal forest. The tree density has been found 769.23 stem ha-1 in teak plantation, 742.86 in sal forest and 671.05 in miscellaneous forest. The basal cover was observed  $1260.75~m^2~h^{-1}$  in miscellaneous forest, 3533.61 in sal forest and 4933.97 in teak plantation. The diversity indexes (Shannon & Fisher-alpha) have been found maximum (2.666 & 12.390) in miscellaneous forest, moderate (2.152 & 8.677) in sal forest and low (1.134 & 6.482) in teak plantation. The species heterogeneity has been found maximum (0.862) in miscellaneous forest, moderate (0.785) in sal forest and minimum (0.399) in teak plantation. The tree species richness and the basal area have also been found more in comparison with the previous reports from the area.

\*Corresponding Author: Lal Babu Chaudhary ⊠ dr\_lbchaudhary@rediffmail.com

#### Introduction

India has been considered as one of the 17 megadiversity centers of the word with a wide range of phytogeographical variations. It consisits of about 64 million hectares forest covers out of which 86% is tropical forest comprising 54% dry deciduous, 37% moist deciduous and 9% wet evergreen & semievergreen (Kaul and Sharma, 1971). As a characteristic feature, the tropical forest shows a huge variation in tree species diversity place to place (Pitman et al., 2002). Among the different phytodiverse regions found in the country, the Terai region is one of them existing from Uttarakhand to W. Bengal. It is the transition zone between two ecoclimatic zones, the Gangatic plain towards south and Bhabhar towards north, along with the sub-Himalayan tracts (Tripathi and Singh, 2009). The region has lost majority of its natural forest due to deforestration chiefly for agriculture and lack of sustainable forest management in last many centuries (Bajpai et al., 2012). Now the natural forests of the region have been restricted to the wildlife protected areas only. Katerniaghat Wildlife Sanctuary (KWS) is also one of the protected areas in this region in Bahraich district of Uttar Pradesh. In earlier the forest has been classified into two major forest types, (i) Sal Forest and (ii) Miscellaneous Forest (Champion and Seth, 1968). Tripathi and Singh, (2009) have also grouped the forest into (i) Natural Forest and (ii) Plantation Forest. Hence, to know the current status of the forest types in the sanctuary area the present study has been carried out.

In this recongregation study only tree species have been considered because the trees are more vulnerable and at the same time the forest ecosystems have always been defined by its tree species diversity (Hubbell and Foster, 1992; Cannon et al., 1998; Rennolls and Laumonier, 2000). Such type of quantitative assessment of the forest tree species provides important information related with the species diversity, distribution, composition as well as forest management for conservation and

sustanaible utilization (Gentry, 1990; Hartshorn, 1990; Phillips et al., 2003; Reddy et al., 2008).

The study area has been sporadically explored in past by various workers (Panigrahi et al., 1969; Saini, 2005; Chauhan et al., 2008; Tripathi et al., 2009; Maliya and Dutt, 2010; Kumar et al., 2011; Maliya, 2012). The preliminary studies on distributional pattern of the tree species of the study area applying ecological parameters including less number of species than expected numbers was carried out by Tripathi and Singh, (2009) and they use only the dominant tree species for the congregation of forest, however, in the present study the data dealing with the large number of species from more areas have been generated to assess their distributional and association pattern. Cluster and PCA analyses have been done on the basis of IVI of the species for the re congregation of forest types in the sanctuary area. Tree density classes have been also made to categorise the tree species within the dominant to very rare class.

The main objectives of the study were (a) analysis of the community structure for species richness, evenness, density and dominance and (b) stratification of the forest tree species into different forest types. The result from such type of studies can be used in future for the assessment of the changes in the environment of the area and their impact on phytodiversity.

# Materials and methods

Study area

The study area Katerniaghat Wildlife Sanctuary (KWS) is situated in Bahraich dstrict of Uttar Pradesh in India (Fig. 1). It lies along Indo-Nepal international boarder and is situated between 27° 41′ - 27° 56′ N and 81° 48′ - 81° 56′ E covering an area of 440 km² with 116 to 165 m elevation. The sanctuary comes under the tropical moist deciduous forest of the Himalayan Terai-Bhabar region (Champion and Seth, 1968; Rodgers and Panwar, 1988). The forest of the sanctuary area has been classified into two major forest types (i) The Sal forest and (ii) The

miscellaneous forest (Champion and Seth, 1968). Pedologicaly the study area is made up of the alluvial soil of the Kaudiyala and Saryu rivers and its tributaries flowing adjoining to it. Geologicaly the sanctuary area has been divided into high and low land areas.

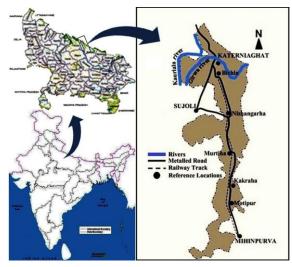


Fig. 1. Location map of study site (Katerniaghat Wildlife Sanctury).

### Climate

A typical tropical monsoonal climate with three distinct seasons i.e. summer (April to June), winter (November to February) and warm-rainy (July to September) prevails in the study area. March and October are considered as transition months between the seasons. The mean maximum temperature ranges from 22°C in January to 40°C in May and the mean minimum temperature ranges from 8°C in January to 27°C in June. The annual rainfall ranges from 36 to 142 cm in winter, 34 to 662 cm in summer and 1294 to 1689 cm in warmrainy seasons (Fig. 2) (Bajpai et al., 2012).

# **Vegetation**

The heterogeneous vegetation of the sanctuary area is tropical moist deciduous type and may be divided into upper stratum, lower stratum and ground vegetation. The upper stratum comprises trees like Shorea robusta, Tectona grandis, Terminalia elliptica, Madhuca longifolia var. latifolia, Ficus benghalensis, Ficus racemosa, Bombax ceiba, Sterculia villosa, Lannea coromandelica etc. which

receives the full solar radiation. The lower stratum is represented by Hymenodictyon orixense, Syzygium cumini, Ehretia laevis, Lagerstroemia parviflora, Diospyros exculpta, Mallotus philippensis, Mallotus nudiflorus, Ficus hispida, Streblus asper etc. The ground vegetation chiefly consists of Bergera koenigii, Glycosmis pentaphylla, Clerodendrum viscosum, Grewia hirsuta, Lantana camara etc. where the minimal light is available.

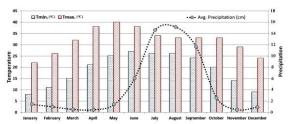


Fig. 2. Climatic variation in Katerniaghat Wildlife Sanctury

# Field data collection and analysis

Random stratified sampling (Greig-Smith, 1983; Krebs, 1989) from 47 quadrates was done for the ecological data collection. Quadrates of 50 x 50 m were plotted to measure the frequency, density, abundance and species dominance (Curtis and McIntosh, 1950). Frequency, density, abundance & species dominance have been used to calculate the species Importance Value Index (IVI) and different diversity indexes (Simpson, 1949; Cootam and Curtis, 1956; Gauch, 1982; ter Braak and Prentice, 1988). Grouping of all random plots has been done by cluster analysis (Ludwig and Reynolds, 1988; Jongman et al., 1995; Rai et al., 2012) using IVI data, employing Bray-Curtis similarity measure and unweighted pair-group moving average (UPGMA) algorithm (Rai et al., 2012). PCA analysis has also been done to find out the differnt groups or communities of the forest area. Cluster and PCA analyses has been performed using multivar option in PAST version 2.12 (Hammer et al., 2001; Hall, 2005; Rai et al., 2012).

#### **Results**

Forest community structure

The cluster and PCA analyses have been conceded by using plot wise IVI data of all random plots to make the natural groups (Table 1). Cluster analysis formed three major clusters (I, II & II) and seven subclusters (A, B, C, D, E, F & G) employing Bray-Curtis similarity measure (Fig. 3). The cluster-I grouped the random plots having Shorea robusta as dominant species (Avg. IVI-122.2), Mallotus philippensis as 1st co-dominant species (Avg. IVI-54.3) and Syzygium cumini as 2nd co-dominant species (Avg. IVI-24.1); cluster-II consists of Mallotus philippensis as dominant species (Avg. IVI-98.2), Mallotus nudiflorus as 1st co-dominant species (Avg. IVI-39.6) and Ficus hispida as 2nd co-dominant species (Avg. IVI-23.0) and cluster-III comprises Tectona grandis as dominant species (Avg. IVI-231.5), Mallotus philippensis as 1st co-dominant species (Avg. IVI-21.0) and Shorea robusta as 2nd co-dominant species (Avg. IVI-7.1) (Table 2). On the basis of the major constituents, the cluster-I termed as the 'Sal Forest',

cluster-II as the 'Miscellaneous Forest' and cluster-III as the 'Teak Plantation'. Almost similar results have been achieved in the PCA analysis except some variations (Fig. 4).

One plot (plot 14) has been found out-grouped in clustering and it has been placed between cluster-I (sal forests) and cluster-III (teak plantation) in PCA. It has *Shorea robusta* (IVI-125.6) as the dominant species with *Tectona grandis* (Avg. IVI-115.2) as 1<sup>st</sup> co-dominant species and *Mallotus philippensis* (Avg. IVI-24.9) as 2<sup>nd</sup> co-dominant species. The presence of *Shorea robusta* as well as *Tectona grandis* with more than 100 IVI is the reason of its placement between the clusters of sal forest and teak plantation in PCA plot.

Table 1. IVI of tree species in three forest types in Katerniaghat Wildlife Sanctury.

Name of Species ↓	IVI →	Miscellaneous Forest	Sal Forest	Teak Plantation	
Acacia catechu		18.3	0	0	
Aegle marmelos		0.0	0	3.8	
Albizia lebbeck		0.9	0	0.0	
Albizia procera		2.8	0.6	0.0	
Alstonia scholaris		0.4	0.0	0.0	
Barringtonia acutangula		1.8	0.0	0.0	
Bauhinia purpurea		0.5	0.0	0.0	
Bombax ceiba		11.0	0.0	0.0	
Bridelia retusa		3.9	2.1	0.7	
Buchanania cochinchinensis		0.0	0.7	0.9	
Acacia concinna var. rugata		1.0	0.0	0.0	
Cassia fistula		0.5	0.7	0.0	
Cordia dichotoma		2.1	0.0	0.0	
Dalbergia sissoo		4.4	0.0	0.0	
Dillenia pentagyna		0.0	2.1	0.0	
Diospyros exculpta		0.0	3.2	3.0	
Ehretia laevis		11.7	3.4	2.3	
Phyllanthus emblica		0.0	0.6	0.0	
Eucalyptus tereticornis		0.0	0.0	0.7	
Ficus bengalensis		0.0	0.0	1.5	
Ficus hispida		23.0	0.0	1.3	
Ficus palmata		0.5	0.0	0.6	
Ficus racemosa		9.7	0.0	1.9	
Ficus rumphii		0.0	1.3	0.0	
Ficus squamosa		5.2	0.0	1.2	
Grewia tiliifolia		5.3	0.0	0.0	
Haldina cordifolia		4.7	0.5	0.0	
Holoptelea integrifolia		0.4	0.0	1.3	
Hymenodictyon orixense		1.1	1.2	0.0	

Kydia calycina	0.9	1.9	0.7
Lagerstroemia parviflora	1.5	12.7	4.6
Lannea coromandelica	12.8	1.3	0.7
Leucaena leucocephala	0.0	1.3	0.0
Litsea glutinosa	0.6	0.6	0.9
Litsea monopetala	0.0	1.8	0.0
Madhuca longifolia var. latifolia	0.0	12.7	1.0
Mallotus nudiflorus	29.6	0.0	0.9
Mallotus philippensis	98.2	54.3	21.0
Miliusa velutina	3.0	10.0	1.6
Mitragyna parvifolia	3.4	0.0	3.9
Bergera koenigii	0.5	0.0	0.0
Desmodium oojeinense	0.0	4.7	0.0
Putranjiva roxburghii	3.0	0.6	0.0
Schleichera oleosa	1.5	4.3	4.4
Semecarpus anacardium	0.0	3.2	0.0
Shorea robusta	10.3	122.4	7.1
Sterculia villosa	0.5	0.0	0.0
Stereospermum chelonoides	0.0	1.2	0.0
Streblus asper	7.9	0.6	0.0
Syzygium cumini	12.9	24.0	2.7
Syzygium heyneanum	2.4	0.0	0.0
Tectona grandis	0.0	8.0	231.5
Terminalia elliptica	0.5	18.3	0
Toona ciliata	0.9	0	0
Ziziphus mauritiana	0.5	0	0
Total	300.0	300.0	300.0

Table 2. Dominant, 1st Co-dominant and 2nd Co-dominant trees species in three forest types with average, standard deviation and minimum - maximum IVI values.

	Avg. IVI	STDEV	Min Max.	Sub Cluster		Avg. IVI	STDEV	Min Max.
Dominant Species					Dominant Species			
Dominant Species					Shorea robusta	66.7	18.5	40.9 - 84.2
Charag rabusta	61 40.9 -		Mallotus philippesis	66.1	30.0	7.8 - 82.2		
Sitorearobasta	122.4	02.2	200.5	A	1 <sup>st</sup> Co-dominant Species			
1st Co-dominant Spe	ecies				Madhuca longifolia var. latifolia	33.2	34.7	11.3 - 89.9
1					2 <sup>nd</sup> Co-dominant Species			
Mallotue philippacie	54.9	27 Q	7.8 -		Lagerstroemia parviflora	27.5	20.8	18.8 - 55.1
Mullotus philippesis 54.3	2/.0	82.2		Dominant Species				
and Co dominant Species				Shorea robusta	169.0	42.9	94.6 - 218.6	
2 Co-dominant species				Species				
Syzygium cumini	24.1	34.0	11.4 -	В	philippesis	40.9	25.4	10.2 - 74.2
			94.0			40.7	37.4	16.7 - 94
					elliptica	10.1	9.3	9.7 - 22.5
				Dominant		•		
Dominant Specie			C			ı	73.8 -	
				philippesis	128.1	33.5	73.8 - 180.3	
Mallotus philippesis	98.2	53.2	14.4 - 180 3		1st Co-dominant			
	Mallotus philippesis  2nd Co-dominant Sp  Syzygium cumini  Dominant Specie	Dominant Species  Shorea robusta 122.4  1st Co-dominant Species  Mallotus philippesis 54.3  2nd Co-dominant Species  Syzygium cumini 24.1  Dominant Specie	Dominant Species  Shorea robusta 122.4 62.2  1st Co-dominant Species  Mallotus philippesis 54.3 27.8  2nd Co-dominant Species  Syzygium cumini 24.1 34.0  Dominant Specie	Dominant Species  Shorea robusta 122.4 62.2 40.9 - 200.5  1st Co-dominant Species  Mallotus philippesis 54.3 27.8 7.8 - 82.2  2nd Co-dominant Species  Syzygium cumini 24.1 34.0 11.4 - 94.0  Dominant Specie	Dominant Species  Shorea robusta  122.4 62.2 40.9 - 200.5  A  1st Co-dominant Species  Mallotus philippesis 54.3 27.8 7.8 - 82.2  2nd Co-dominant Species  Syzygium cumini 24.1 34.0 11.4 - 94.0  Dominant Specie  Mallotus philippesis 08.2 52.2 14.4 -	Dominant Species  Shorea robusta  122.4   62.2   40.9 - 200.5   A	Nominant Species	Name

	1	1 1	Ī	İ	Í	Syzygium cumini	17.0	23.4	7.6 - 75.5
				l		2 <sup>nd</sup> Co-dominant	-/	_0.1	70.0
	1st Co-dominant Sp	ecies				Species			1
	- *** ********************************					Lannea coromandelica	15.3	27.3	12 - 69.5
				8.4 -		Shorea robusta	15.0	27.5	12.8 - 83.1
	Mallotus nudiflorus	29.6	49.9	176.2		Dominant Species			1 00.1
					-	Mallotus	75.1	68.9	28.2 -
	2 <sup>nd</sup> Co-dominant Sp	oecies				nudiflorus  1st Co-dominant Species	, ,		176.2
				7.4 -	D	Ficus hispida	48.6	61.7	65.3 - 154
	Ficus hispida	23.0	41.2	154.1		2 <sup>nd</sup> Co-dominant Species			1 -01
						Mallotus philippesis	33.6	12.8	14.4 - 47.7
	Dominant Species					Dominant Species			17.7
	Dominant Species					Tectona grandis	90.7	0.5	90.3 - 91.0
				90.3 -		1 <sup>st</sup> Co-dominant Species			
	Tectona grandis 231.5 72.3 90.3 300.0	E	Mallotus philippesis	39.7	42.9	9.4 - 70.0			
						2 <sup>nd</sup> Co-dominant Species			II.
	1st Co-dominant Sp	ecies				Mitragyna parvifolia	25.2	15.1	14.5 - 35.8
				7.2 -		Dominant Species			1 33.0
	Mallotus philippesis	21.0	24.3	70.0		Tectona grandis	282.5	19.0	252.8 - 300.0
				•		1 <sup>st</sup> Co-dominant Species			
III	2 <sup>nd</sup> Co-dominant Sp	ecies			F	Mallotus philippesis	2.8	5.1	7.2 - 12.6
				12.8 -		Ficus bengalensis	2.8	7.3	19.4
	Shorea robusta	7.1	18.4	66.1		2 <sup>nd</sup> Co-dominant Species			
						Lagerstroemia parviflora	2.6	6.9	18.3
						Dominant Species			II.
						Tectona grandis	212.7	21.0	183.3 - 232.8
						1 <sup>st</sup> Co-dominant Species			
					G	Mallotus philippesis	43.4	5.4	36.1 -
						2 <sup>nd</sup> Co-dominant Species			47.6
					>00010C				

 Table 3. Inventory details of tree species in Katerniaghat Wildlife Sanctuary.

Variable	Forest Types						
	<b>Miscellaneous Forest</b>	Sal Forest	Teak Plantation				
Dominant species	Mallotus philippensis	Shorea robusta	Tectona grandis				
Number of species	40.0	31.0	25.0				
Number of genera	33.0	29.0	20.0				
Number of families	24.0	21.0	18.0				
Site specific species	15.0	9.0	3.0				

g : / ::			
Species/genus ratio	1.21	1.07	1.25
Species richness (spp. ha <sup>-1</sup> )	52.63	55.35	48.07
Tree Density (stem ha <sup>-1</sup> )	671.05	742.86	769.23
Basal cover (m² ha-1)	22.97	65.90	89.07
Different indexes			
Dominance_D	0.138	0.215	0.601
Simpson_1-D	0.862	0.785	0.399
Fisher_alpha	12.390	8.677	6.482
Shannon_H	2.666	2.152	1.134
Equitability_J	0.723	0.627	0.352
Evenness_e^H/S	0.360	0.278	0.124

Table 4. Density classes of tree species in Katerniaghat Wildlife Sanctuary.

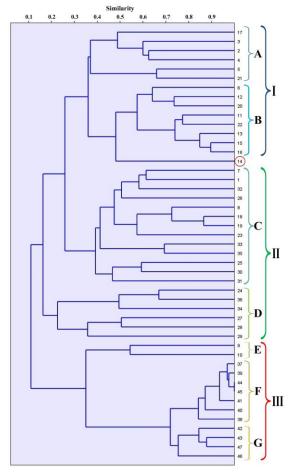
	Number of Species						
Density Classes	10- 500	60- 20	21 - 50	51 - 100	> 100		
Density Classes	Very rare	Rare	Common	Prominent	Dominant		
Sal Forest	18	5	5	1	2		
Miscellaneous Forest	19	14	4	2	1		
Teak plantation	17	6	0	1	1		

The paired group Bray-Curtis clustering placed the random plot 5 in the cluster-I while PCA analysis grouped it into cluster-II. All over species composition of the plot 5 and presence of Terminalia elliptica (Avg. IVI-87.3) as 1st co-dominant species is responsible for it's placement within cluster-I while the presence of 37 individuals of dominant species i.e. Mallotus philippensis with higher Avg. IVI-133.1 (as compare to cluster-I) and presence of only 5 individuals of 2<sup>nd</sup> co-dominant species i.e. Shorea robusta with lesser IVI-28.2 (as compare to Shorea robusta in cluster-I) make it closer to the subcluster-C of cluster-II which has the Mallotus philippensis with Avg. IVI-128.1 (very close to plot 5) as dominant species and Shorea robusta with Avg. IVI-15.0 as 2<sup>nd</sup> co-dominant species.

Sub-cluster-A of cluster-I has been drown closer to sub-cluster-C of cluster-II in the PCA plot, because sub-cluster-A grouped the random plots having Shorea robusta (Avg. IVI-66.7) as well as Mallotus philippensis (Avg. IVI-66.1) as dominant species which is closer to cluster-II where Mallotus philippensis is also a dominant species (Avg. IVI-98.2). Out of two sub-clusters C & D in cluster-II,

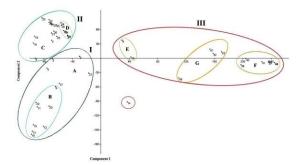
sub-cluster-A of cluster-I shows its closeness with sub-cluster-C due to the presence of Shorea robusta in random plots of both the sub-clusters (as dominant species in sub-cluster-A with Avg. IVI-66.7 and as 2<sup>nd</sup> co-dominant species in sub-cluster-C with Avg. IVI-15.0). While sub-cluster-B of cluster-I has been drown away from the cluster-II, because it grouped the random plots having Shorea robusta as dominant species (with higher Avg. IVI-169.0) which is about four times higher than the 1st co-dominant species i.e. Mallotus philippensis & Syzygium cumini (Avg. IVI-40.9 & 40.7). Due to the presence of Shorea robusta with highest Avg. IVI in cluster-B, it can be named as "Sal Dominated Forest", while cluster-B can be named as "Sal Miscellaneous Forest" due to the mixed vegetation dominated by Shorea robusta (Avg. IVI-66.7) & Mallotus philippensis (Avg. IVI-66.1).

The cluster-II has been further grouped in two subclusters C & D. In the PCA plot both the sub-clusters can not be differentiated easily but in the cluster graph these both clusters can be easily differentiated. The major reason behind this regrouping of subclusters C and D is the species composition, which is completely different. Sub-cluster-C has been represented by the Mallotus philippensis (Avg. IVI-128.1) as dominant species, Syygium cumini (Avg. IVI-17.0) as 1st co-dominant species and Lannea coromandelica (Avg. IVI-15.3) & Shorea robusta (Avg. IVI-15.0) as 2nd co-dominant species. Due to the presence of Mallotus philippensis with more than 100 IVI value (seven to eight times higher than its 1st & 2nd co-dominant species); cluster-C can be named as "Mallotus Miscellaneous Forest". Subcluster-D has been represented by Mallotus nudiflorus (Avg. IVI-75.1) as dominant species and Ficus hispida (Avg. IVI-48.6) as 1st co-dominant species. Both the species have been considered as the moisture adoring species, which grow luxuriantly in the low land areas. The presence of these low land species in this sub-cluster, it can be named as "Low Land Miscellaneous Forest".

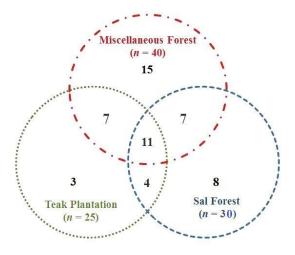


**Fig. 3.** Cluster showing the different forest types: I-Sal Forest (A- Sal Dominated Forest and B- Sal Miscellaneous Forest), II- Miscellaneous Forest (C-

Mallotus Miscellaneous Forest and D- Low Land Miscellaneous Forest) and III-Teak Plantation (E-Teak Miscellaneous Forest, F- Pure Teak Plantation and G- Teak Sal Miscellaneous Forest) with one out group (Plot-14).



**Fig. 4.** PCA plot showing the different forest types: I-Sal Forest (A- Sal Dominated Forest and B- Sal Miscellaneous Forest), II- Miscellaneous Forest (C-*Mallotus* Miscellaneous Forest and D- Low Land Miscellaneous Forest) and III-Teak Plantation (E-Teak Miscellaneous Forest, F- Pure Teak Plantation and G- Teak Sal Miscellaneous Forest) with one out group (Plot-14).



**Fig. 5.** Venn diagram showing species distribution in different forest types of Katerniaghat Wildlife Sanctury.

Within the cluster-III three sub-clusters E, F and G are easily visible in PCA plot as well as in cluster graph. The species composition in the sub-cluster-E moves towards the cluster-II, due to the presence of *Mallotus philippensis* (Avg. IVI-39.7) as 1<sup>st</sup> codominant species and *Mitragyna parvifolia* (Avg. IVI-25.2), *Aegle marmelos* (Avg. IVI-24.8),

Schleichera oleosa (Avg. IVI-23.9) & Lagerstroemia parviflora (Avg. IVI-20.6) as 2nd co-dominant species. The presence of these miscellaneous species with Tectona grandis (Avg. IVI-90.7) may be used to name this sub-cluster as "Teak Miscellaneous Forest". The sub-cluster-F has been ploted far away from sub-cluster-E. The random plots of this subcluster have been dominated by Tectona grandis with an Avg. IVI-282.5. The IVI of Tectona grandis is very high here, due to the monoculture of the species. Thus this sub-cluster may be named as "Pure Teak Plantation". The sub-cluster-G has been placed in between sub-cluster E & F. The random plots of this sub-cluster have been dominated by Tectona grandis (Avg. IVI-212.7) with Mallotus phileppensis (Avg. IVI-43.4) as 1st co-dominant species and Shorea robusta (Avg. IVI-16.5) as 2nd codominant species. The lower Avg. IVI value of Tectona grandis (as compare to sub-cluster-F) and presence of Shorea robusta as 2nd co-dominant species is responsible for its place in the PCA plot. Thus this sub-cluster may be named as "Teak Sal Miscellaneous Forest".

Tree species richness, dominance, density and heterogeneity

The number of species has been found maximum in miscellaneous forest (40 spp.) followed by sal forest (30 spp.) and teak plantation (25 spp.) (Table 3). The species/genus ratio (Ricklefs and Miller 2000) has been found maximum in teak plantation showing its recent diversification. In all the three forest types 11 tree species have been found commonly growing (Fig. 5). The miscellaneous forest shows maximum tree diversity by having 15 site specific species which are present here only (i.e. Acacia catechu, Albizzia lebbek, Alstonia scholaris, Barringtonia acutangula, Bauhinia purpurea, Bombax ceiba, Acacia concinna var. rugata, Cordia dichotoma, Dalbergia sissoo, Grewia teliafolia, Bergera koenigii, Sterculia vilosa, Syzygium heyneanum, Toona ciliata and Zizyphus mauritiana), while the sal forest and teak plantation have only 8 species (i.e. Dillenia penatagyna, Phyllanthus emblica, Ficus rumphii, Leucaena lucosifela, Litsea monopetala, Desmodium oojeinense, Semecarpus anacardium and Stereospermum suaveolens) and 3 species (i.e. Aegle marmelos, Eucalyptus tereticornis and Ficus bengalensis) as site specific respectively (Fig. 5).

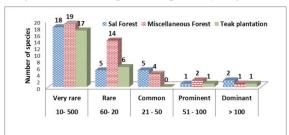


Fig. 6. Species dominance and rarity Katerniaghat Wildlife Sanctury.

Although the maximum number of species has been reported from the miscellaneous forest but the maximum species richness (i.e. 55.35 species ha-1) has been calculated in the sal forest followed by the miscellaneous forest (52.63). The teak plantation shows the minimum species richness (48.07) because of its monodominant nature. When we talk about the tree density (stem ha-1), the maximum value has been found in teak plantation (769.23) due to sufficient distance provided between the trees. The miscellaneous forest having the highest number of species shows the minimum (671.05) tree density. The sal forest represents the moderate (742.86) tree density in the sanctuary area. As similar to the tree density, the maximum values of basal cover has been observed in the teak plantation (89.07 m<sup>2</sup> ha<sup>-1</sup>) and the minimum in miscellaneous forest (22.97). The sal forest represents the moderate (65.90) basal cover.

The tree dominance has been computed maximum (0.601) for the teak plantation which clearly indicates its monoculture nature having only Tectona grandis as dominant tree species and it has been also supported by the lowest Simpson Index value (0.399). The minimum dominance has been computed for the miscellaneous forest (0.138) having the maximum number of dominant tree species i.e. highest heterogeneity (Simpson Index-0.862). The sal forest shows the moderate heterogeneity of tree species having the dominance value 0.215 and Simpson Index 0.785.

The highest Shannon diversity index (2.666) as well as Fisher alpha diversity index (12.390) has been found in the miscellaneous forest, which indicates its highest species diversity in this forest type because the higher value of these two diversity indexes denote the higher species diversity. In the teak plantation the diversity indexes have been computed minimum in both Shannon diversity (1.134) and Fisher alpha diversity (6.482) which reveals its lowest tree diversity among the three forest types. In the sal forest, Shannon and Fisher alpha diversity indexes have been computed 2.152 and 8.677 respectively representing its moderate tree diversity in the sanctuary area.

The value of Equitability (0.723) and Evenness (0.360) has been computed maximum for the miscellaneous forest where the species are more evenly distributed in random quadrates while in the teak plantation the equitability and evenness have been found minimum 0.124 and 0.352 respectively. The sal forest represents the moderate value of the Equitability and Evenness (0.360 and 0.723 respectively).

# Tree density classes

The five density classes have been made by using the tree density (stem ha-1). Here it has been tried to list the number of dominant (> 100 stem ha-1), prominent (51-100 stem ha-1), common (21-50 stem ha-1), rare (6-20 stem ha-1) and very rare (1-5 stem ha-1) tree species in different forest types of the forest in Katerniaghat Wildlife Sanctuary (Table 4). In sal forest 2, 1, 5, 5 and 18 tree species comes under dominant, prominent, common, rare and very rare density classes respectively while in the case of miscellaneous forest the number have been changed to 1, 2, 4, 14 and 19 respectively (Fig. 6). In the teak plantation single tree species has come in dominant and prominent classes, 6 in rare, 17 in very rare and no species has come in common class.

In sal forest Shorea robusta and Mallotus philippensis come under dominant density class, having 275.00 and 158.93 tree density respectively out of total tree density (742.86). Syzygium cumini with 53.57 tree density is the only prominant species according to density classes. Terminalia elliptica (33.93), Lagerstromia parviflora (32.14), Madhuca longifolia var. latifolia (32.14), Miliusa velutina (32.14) and Tectona grandis (25.00) come under common density class. Desmodium oojeinense (10.71), Schleichera oleosa (10.71), Ehretia laevis (8.93), Diospyros exculpta (7.14) and Samecarpus anacardium (7.14) belong to rare density class. Abutilaon indicum (5.36), Bridelia retusa (5.36), Dillenia pentagyna (5.36), Litsea monopetala (5.36), Ficus rumphii (3.57), Hymenodictyon orixense (3.57), Lannea coromandelica (3.57), lucosifelia Stereospermum Leucaena (3.57),suaveolens (3.57),Albizzia prosera (1.79),Buchanania cochinchinensis (1.79), Casia fistula (1.79), Diospyros exculpta (1.79), Phyllanthus emblica (1.79), Haldina cardifolia (1.79), Litsea glutinosa (1.79), Putranjiva roxburbhii (1.79) and Streblus asper (1.79) fall within very rare density class.

In miscellaneous forest only Mallotus philippensis comes under dominant density class with 230.26 tree density out of 671.05 total tree densities. Ficus hispida (59.21) and Mallotus nudiflorus (52.63) are the prominant tree species. Acacia catechu (39.47), Ehretia laevis (30.26), Lannea coromandelica (30.26) and Syzygium cumini (27.63) are common. Streblus asper (19.74), Shorea robusta (17.11), Bombax ceiba (15.79), Ficus racemosa (14.47), Grewia teliafolia (14.47), Ficus saemocarpa (11.84), Haldina cardifolia (11.84), Bridelia retusa (9.21), Miliusa velutina (9.21), Putranjiva roxburbhii (9.21), Albizzia prosera (7.89), Dalbergia sissoo Mitragyna parvifolia Barringtonia acutangula (6.58) come under rare category. Cordia dichotoma (5.26), Lagerstromia parviflora (3.95), Albizzia lebbek (2.63), Acacia

concinna var. rugata (2.63), Hymenodictyon orixense (2.63), Schleichera oleosa (2.63), Syzygium heyneanum (2.63), Abutilon indicum (1.32), Alstonia scholaris (1.32), Bauhinia purpurea (1.32), Cassia fistula (1.32), Ficus palmata (1.32), Holoptelea integrifolia (1.32), Litsea glutinosa (1.32), Bergera koenigii (1.32), Sterculia vilosa (1.32), Terminalia elliptica (1.32), Toona ciliate (1.32) and Zizyphus mauritiana (1.32) belong to very rare group.

In teak plantation only Tectona grandis comes under dominant density class with 598.08 density out of 769.23 total tree density. Mallotus philippensis (59.62) is the only prominant tree species. Lagerstromia parviflora (13.46), Shorea robusta (13.46), Aegle marmelos (9.62), Schleichera oleosa (9.62), Mitragyna parvifolia (7.69) and Syzygium cumini (7.69) come under rare category. In this forest type, not a single species comes under common density class. Diospyros exculpta (5.77), Ehretia laevis (5.77), Miliusa velutina (5.77), Ficus hispida (3.85), Ficus racemosa (3.85), Ficus saemocarpa (3.85), Abutilon indicum (1.92), Bridelia retusa (1.92), Buchanania cochinchinensis Eucalyptus tereticornis (1.92), Ficus bengalensis (1.92), Ficus palmata (1.92), Holoptelea integrifolia (1.92), Lannea coromandelica (1.92), Litsea glutinosa (1.92), Madhuca longifolia var. latifolia (1.92) and Mallotus nudiflorus (1.92) come under the very rare class.

#### **Discussion**

The results of PCA and Cluster analyses clearly formed three major forest types (sal forest, miscellaneous forest and teak plantation) on the basis of IVI value of the species in the random plotes. In teak plantation, 50 to 60 years before Tectona grandis was introduced by the forest department in the open land for timber production (Tripathi and Singh, 2009), however, gradually many tree species have been occupied the good position within this monodominant plantation (Table 2). The sal forest is the oldest natural forest of the Terai region dominated by Shorea robusta. However, the miscellaneous forest can be consider as the developing forest type since it consists of maximum site specific species and common species. The common species growing in this forest type after reaching here via immigration from nearby dissimilar habitats which have the higher ecological amplitude (Shimada and Wilson, 1985). This forest type is very important for the diversity point of view due to the presence of maximum site specific species with low ecological amplitude and less capable to survive outside their specific habitat.

In the Venn diagram (Fig. 5) some of the species have been found confined within a particular forest type. These species can be termed as habitat specific species (Varghese and Menon, 1999). These species have been considered valuable for the conservation point of view, because they have low ecological amplitude which may be responsible for its narrow distribution (Prasad et al., 2007) and have more risk to be vanished. On the other hands some of the species have been found commonly distributed throughout the forest in all kinds of habitats. The reason behind the existence of these companion species in all forest types may be the overlapping of nich requirements (Prasad et al., 2007).

To compare the species richness, tree density and basal cover of the three forest types per hectare scale has been taken (Table 3). The mean species richness has been found 52.02 species ha-1, ranging from 48.07 in teak plantation to 55.35 in sal forest. These values of species richness have been found within the range of tropics i.e. 20-307 spp. ha-1 (Campbell et al., 1992; Valencia et al., 1994). These values of the species richness have been found lower in comparison with the humid tropical evergreen forest (61 species ha-1) (Tripathi et al., 2004), but higher than the tropical rain forest (43 species ha-1) (Strasberg, 1996) and tropical moist forest (45 species ha<sup>-1</sup>) (Tripathi, 2001). The species richness in the study area has been found higher than the tropical dry forests of Mirzapur (9-14) (Singh and Singh, 1991), Similipal Biosphare Reserve (19-36)

(Mishra et al., 2008) and tropical dry evergreen forest of Tanil Nadu (19-35) (Venkateshwaran and Parthasarathy, 2003; Mani and Parthasarathy, 2005) but lower than tropical dry deciduous forest of Andhra Pradesh (69) (Reddy et al., 2008) and tropical wet evergreen forest of Kalakad (80-85) (Parthasarathy, 1999). More significantly, the mean species richness has been found greater than the earlier report from the area made by Tripathi and Singh, (2009) in which they have recorded only 44 tree species in 7.02 hectares.

The mean tree density has been found 727.71 stem ha-1, ranging from 671.05 in miscellaneous forest to 769.23 in teak plantation. These values of tree density have been found within the range of tropics i.e. 276-935 stem ha-1 (Murali et al., 1996; Sundarapandian and Swamy, 1997; Ghate et al., 1998; Mani and Parthasarathy, 2005). The tree density in the sanctuary area has been found higher than the tropical evergreen forests of Western as well as Eastern Ghats where it ranges from 419-716 stem ha-1 (Singh et al., 1984; Ganesh et al., 1996; Ghate et al., 1998; Parthasarathy, 1999; Chittibabu and Parthasarathy, 2000) and tropical deciduous forests (150-627 stem ha-1) (Jha and Singh, 1990; Singh and Singh, 1991; Varghese and Menon, 1998; Shrestha and Jha, 1997; Pandey and Shukla, 2003; Rautiainen, 1999). Tropical moist forest of Singapore (604 stem ha<sup>-1</sup>) (Swan Jr, 1988) and tropical rain forest of Costa Rica (391-617 stem ha-1) (Heaney and Proctor, 1990) also have the lower tree density than that of the sanctuary area. Some tropical deciduous forest of Eastern Ghats ranging from 735-810 stem ha-1 (Mishra et al., 2008; Reddy et al., 2008), tropical evergreen forest of Eastern Ghats and North-East India ranging from 750-935 stem ha-1 (Visalakshi, 1995; Mani and Parthasarathy, 2005; Supriya and Yadava, 2006) and tropical rain forest of Amazon and Malaysia ranging from 1054-1420 stem ha-1 (Campbell et al., 1986; Proctor et al., 1988) show the higher tree density than the sanctuary area.

The mean basal area of tree species in KWS has been found 59.31 m2 ha-1, ranging from 22.97 in miscellaneous forest to 89.07 in teak plantation. These values of basal area have been found within the range of tropics i.e. 7-104 m<sup>2</sup> ha<sup>-1</sup> (Singh et al., 1984; Jha and Singh, 1990; Mishra et al., 2008). The basal area has been found higher than the tropical evergreen forests ranging from 11-82.76 m<sup>2</sup> ha<sup>-1</sup> (Campbell et al., 1992; Visalakshi, 1995; Strasberg, 1996; Chittibabu and Parthasarathy, 2000; Mani and Parthasarathy, 2005; Supriya and Yadava, 2006) and tropical deciduous forest ranging from 7-61 m<sup>2</sup> ha-1 (Jha and Singh, 1990; Singh and Singh, 1991; Singh et al., 1995; Shrestha and Jha, 1997; Varghese and Menon, 1998; Singh et al., 2005; Tripathi and Singh, 2009) in India. The basal area of tropical rain forest of Malaysia i.e. 26-46 m2 ha-1 (Proctor et al., 1988) also has the lower basal area than the sanctuary area. The basal cover of some evergreen forests of Western Ghat ranging from 29-103 m2 ha-1 (Singh et al., 1984; Ganesh et al., 1996; Parthasarathy, 1999), Similipal Biosphere Reserve  $48.7-10.9 \text{ m}^2 \text{ ha}^{-1}$  (Mishra et al., 2008) and the tropical rain forest of Amazonia 28 to 68 m<sup>2</sup> ha<sup>-1</sup> (Campbell et al., 1986) have been found higher than the forest of sanctuary area. More significantly, the basal area of the adjacent forests ranging from 16-61 m<sup>2</sup> ha<sup>-1</sup> (Shrestha and Jha, 1997; Singh et al., 2005; Tripathi and Singh, 2009) has been found lower than the basal cover of the study area.

In KWS the mean dominance (0.318) ranges from 0.138 in miscellaneous forest to 0.601 in teak plantation has been found much higher than the average dominance value (0.06) of tropical forests (Knight, 1975). However, it is within the range from Indian tropics (0.21-0.92) repoted Parthasarathy et al., (1992) and Visalakshi, (1995). Supriya and Yadava, (2006) have also reported the dominance value 0.23-0.97 from the tropical semievergreen forest of Manipur. The dominance range of KWS has been found very close to the dominance values of sal and teak forest (0.07-0.63)

of Madhya Pradesh, India (Prasad and Pandey, 1992).

The Shannon diversity index of the three forest types in KWS has been found 1.984 as a mean with a range from 1.134 in teak plantation to 2.666 in miscellaneous forest. The value of diversity index has been found within the range (0.83-4.10) of diversity index in Indian tropics (Singh et al., 1984; Parthasarathy et al., 1992; Visalakshi, 1995). Further, it has been found closer to the diversity indexes of tropical forest of Madhya Pradesh (0.32-3.76) (Prasad and Pandey, 1992) and Kodayar, Western Ghats (2.20-2.65) (Sundarapandian and Swamy, 2000). The higher value of diversity index has been reported from the tropical forest of Kalakad, Western Ghats (3.31-3.69) (Parthasarathy et al., 1992), Silent Valley, Kerala (3.52-4.15) (Singh et al., 1984) and tropical forest of Garo hills (2.47-4.27) (Ashish et al., 2006) in comparison to KWS. This diversity index is very low than the tropical forest of Panama (5.06-5.40) (Knight, 1975) which may due to high degree of disturbance and anthropogenic pressure such as grazing, burning, collection of woods and medicinal etc. (Foster 1990; Jayasingam and plants Vivekanantharaja, 1994).

#### Conclusion

The tree species richness and the basal area of the KWS have been found higher than previous report made by Tripathi and Singh, (2009). This increment in the value of species richness is a very good sign for the biodiversity expansion. The high species richness in the sal forest indicates that this forest stand is capable to achieve its natural conditions, if we reduce the biotic stress. The high species diversity and low basal cover in miscellaneous forest clearly indicate the growing nature of this forest types with maximum number of species having younger individuals. The teak plantation is the recently diversified forest which has been indicated by higher value of species genus ratio (Tripathi and Singh, 2009; Shukla, 2009; Reddy et al., 2007; Reddy et al., 2008; Parthasarathy, 2001; Mani and Parthasarathy,

2005). A very high IVI value (231.5) of Tectona grandis is responsible for the monodominant nature of the teak plantation. Champion and Seth, (1968) have classified the forest of the sanctuary area into two major types (i) Sal Forest and (ii) Miscellaneous Forest, however, Tripathi and Singh, (2009) have categorised them into (i) Natural Forest and (ii) Plantation Forest on the basis of the dominant species and species composition. In the present study three major forest types have been categorised (i) Sal forest, (ii) Miscellaneous forest and (iii) Teak Plantation on the basis of cluster and PCA analyses using IVI data of the random quadrates. The microclimatic condition may be the main cause of this diversification of forest types (Behera et al., 2012).

#### Acknowledgements

The authors are grateful to Dr. C. S. Nautiyal, **CSIR-National Botanical** Research Director, Institute, Lucknow for providing necessary facilities. The financial support was received from DST, New Delhi under GAP-215725. The authors are also thankful to PCCF (Wildlife), Government of Uttar Pradesh, Lucknow, DFO (Wildlife) Bahraich and all forest staff of Katerniaghat Wildlife Sanctuary, Uttar Pradesh for granting permission and logistic support to conduct the study in the sanctuary area.

# References

Kumar A, Bruce GM, Saxena A. 2006. Tree species diversity and distribution patterns in tropical forests of Garo Hills. Current Science 91(10), 25.

Bajpai O, Kumar A, Mishra AK, Sahu N, Behera SK, Chaudhary LB. 2012. Phenological study of two dominant tree species in tropical moist deciduous forest from the Northern India. International Journal of Botany 8(2), 66-72.

Behera SK, Mishra AK, Sahu N, Kumar A, Singh N, Kumar A, Bajpai O, Chaudhary LB, Khare PB, Tuli R. 2012. The study of microclimate in response to different plant community association in tropical moist deciduous forest from northern India. Biodiversity and Conservation 21(5), 1159-1176.

Campbell DG, Daly DC, Prance GT, Maciel UN. 1986. Quantitative ecological inventory of Terra firma and the Varzea tropical forest on the Rio Xingu, Brazilian Amazon. Brittanica 38, 369-393.

Campbell DG, Stone JL, Rosas JrA. 1992. A comparison of the phytosociology and dynamics of three foodplain (Várzea) forests of known ages, Rio Juruá, western Brazilian Amazon. Botanical Journal of Linnean Society 108, 213-237.

Cannon CH, Peart DR, Leighton M. 1998. Tree species diversity in commercially logged Bornean rainforest. Science 28, 1336-1368.

Champion HG, Seth SK. 1968. A Revised Survey of the Forest Types of India. Publication Division, Govt. of India, New Delhi.

Chauhan DS, Dhanai CS, Singh B, Chauhan S, Todaria NP, Khalid MA. 2008. Regeneration and tree diversity in natural and planted forests in a Terai - Bhabhar forest in Katarniaghat Wildlife Sanctuary, India. Tropical Ecology 49(1), 53-67.

Chittibabu CV, Parthasarathy N. Attenuated tree species diversity in human-impacted tropical evergreen forest sites at Kolli hills, Eastern Ghats, India. Biodiversity and Conservation 9, 1493-1519.

Cootam G, Curtis JT. 1956. The use of distance measures in phytosociology sampling. Ecology 37, 451-460.

Curtis JT, McIntosh RP. 1950. The interrelations of certain analytic and synthetic phytosociological characters. Ecology 31(3), 434-455.

Foster RB. 1990. Heterogeneity & disturbance in tropical vegetation. in: Soule, M. E. and Wilcox, B. A. (Eds.) Conservation Biology: An Evolutionary-**Ecological** Perspective. Sinauer Associates, Sunderland, 75-92.

Ganesh T, Ganesan R, Devy MS, Davidar P, Bawa KS. 1996. Assessment of plant biodiversity at a mid-elevation evergreen forest of Kalakad-Mundanthurai Tiger Reserve. Current Science 71, 379-392.

Gauch JrHG. 1983. Multivariate analysis in community structure. Cambridge University Press., Cambridge.

Gentry AH. 1990. Floristic similarities and differences between southern Central America and Upper and Central Amazonia, in: Gentry, A. H. (Ed.) Four neotropical rainforests. Yale University Press., New Haven, USA, 141-157.

Ghate U, Joshi NV, Gadgil M. 1998. On the patterns of tree diversity in the Western Ghats of India. Current Science 75, 594-603.

Greig-Smith P. 1983. Quantitative plant ecology, 3rd edition. Blackwell, London.

Hall A. 2005. The plant community of Bergen swamp, NY, a rich minerotrophic mire, in: The environ-mental gradients and plant communities of Bergen swamp, NY, USA. Thesis Report, Rochester Institute of Technology, New York.

Hammer Ø, Harper DAT, Ryan DP. 2001. PAST: Paleontological statistics software package for education and data analysis. Palaeontol Electron **4(1)**, 9.

Hartshorn GS. 1990. An overview of neotropical forest dynamics, in: Gentry, A. H. (Ed.) Four neotropical rainforests. Yale University Press, New Haven, Connecticut, USA, 585-599.

Heaney A, Proctor J. 1990. Preliminary studies on forest structure and floristics on Volcan Barva, Costa Rica. Journal of Tropical Ecology 6, 307-320.

Hubbell SP, Foster RB. 1992. Short-term dynamics of a Neotropical forest: why ecological research matters to tropical conservation and management. Oikos 63, 48-61.

Jayasingam T, Vivekanatharaja S. 1994. Vegetation survey of the Wasgamuwa National Park, Srilanka. Vegetation 13(1), 1-8.

Jha CS, Singh JS. 1990. Composition and dynamics of dry tropical Forest in relation to soil texture. Journal of Vegetation Science 1, 609-614.

Jongman RHG, ter Braak CJF, van Tongeren OFR. (Eds.) 1995. Data analysis in community and landscapeecology. Cambridge University Press, Cambridge.

Kaul ON, Sharma DC. 1971. Forest type statistics. Indian Forester 97, 432-436.

Knight DH. 1975. A phytosociological analysis of species rich tropical forest on Barro Colorado Island, Panama. Ecological Monographs 45, 259-289.

Krebs CJ. 1989. Ecological methodology. Harper and Row Publication, New York.

Kumar A, Bajpai O, Mishra AK, Sahu N, Behera SK, Chaudhary LB. 2011. Assessment of Diversity in the Genus Ficus L. (Moraceae) of Katerniaghat Wildlife Sanctuary, Uttar Pradesh, India. American Journal of Plant Science 2, 78-92.

Ludwig JA, Reynolds JF. 1988. Statistical ecology. A primer on methods and computing. John Wiley, London.

Maliya SD, Datt B. 2010. A Contribution to the Flora of Katarniyaghat Wildlife Sanctuary, Bahraich District, Uttar Pradesh. Journal of Economic and Taxonomic Botany 34(1), 42-68.

Maliya SD. 2012. Addition to the flora of Katarniaghat Wildlife Sanctuary, Bahraich district, Uttar Pradesh. Journal of Economic and Taxonomic Botany 36(2), 419-426.

Mani S, Parthasarathy N. 2005. Biodiversity assessment of trees in five inland tropical dry evergreen forests of peninsular India. Systematics and Biodiversity 3(1), 1-12.

Mishra RK, Upadhyay VP, Mohanty RC. 2008. Vegetation Ecology of the Similipal Biosphere Reserve, Orissa, India. Applied Ecology and Environmental Research 6(2), 89-99.

Murali KS, Uma S, Shaanker U, Ganeshaiah KN, Bawa KS. 1996. Extraction of forest products in the forests of Biligirirangan Hills, India and impact of NTFP extraction on regeneration, population structure and species composition. Economic Botany 50, 252-269.

Pandey SK, Shukla RP. 2003. Plant diversity in managed sal (Shorea robusta Gaertn.) forests in Gorakhpur, India: species composition, regeneration and conservation. Biodiversity and Conservation 12, 2295-2319.

Panigrahi G, Singh AN, Misra OP. 1969. Contribution to the Botany of the Tarai Forests of the Bahraich District of Uttar Pradesh. Bulletin of Botanical Survey of India 11(1&2), 89-114.

Parthasarathy N. 1999. Tree diversity and distribution in undisturbed and human-impacted sites of tropical wet evergreen forest in southern Western Ghats, India. Biodiversity and Conservation 8, 1365-1381.

Parthasarathy N. 2001. Changes in forest composition and structure in three sites of tropical evergreen forest around Sengaltheri, Western Ghats. Current Science 80(3), 389-393.

Phillips OL, Martinez RV, Vargas PN, Monteagudo AL, Zans MC, Sanchez WG, Cruz AP, Timana M, Yli-Halla M, Rose S. 2003. Efficient plot-based floristic assessment of tropical forests. Journal of Tropical Ecology 19, 629-645.

Pitman NCA, Terborgh JW, Silman MR, Percy NV, Neill DA, Ceron CE, Palacios WA, Aulestia M. 2002. A comparison of tree species diversity in two upper Amazonian forests. Ecology 83, 3210-3224.

Prasad PRC, Reddy CS, Dutt CBS. 2007. Phytodiversity assessment of tropical rainforest of North Andaman Islands, India. Research Journal of forestry 1(1), 27-39.

Prasad R, Pandey RK. 1992. An observation on plant diversity of Sal and Teak forest in relation to intensity of biotic impact of various distances from habitation in Madhya Pradesh: A case study. Journal of Tropical Forestry 8(1), 62-83.

Proctor J, Lee YF, Langley AM, Munro WRC, Nelson T. 1988. Ecological studies on Gunung Silam, a small ultrabasic mountain Sabah, Malaysia. Journal of Ecology **74**, 455-463.

Rai H, Upreti DK, Gupta RK. 2012. Diversity and distribution of terricolous lichens as indicator of habitat heterogeneity and grazing induced trampling a temperate-alpine shrub and meadow. Biodiversity and Conservation 21, 97-113.

Rautiainen O. 1999. Spatial yield model for Shorea robusta in Nepal. Forest Ecology and Management 119, 151-162.

Reddy CS, Babar S, Giriraj A, Reddy KN, Rao KT. 2008. Structure and floristic composition of tree diversity in tropical dry deciduous forest of Estern Ghats, Southern Andhra Pradesh, India. Asian Journal of Scientific Research 1(1), 57-64.

Reddy CS, Pattanaik C, Mohapatra A, Biswal AK. 2007. Phytosociological Observations on Tree Diversity of Tropical Forest of Similipal Biosphere Reserve, Orissa, India. Taiwania 52(4), 352-359.

Rennolls K, Laumonier Y. 2000. Species diversity structure analysis at two sites in the tropical rain forest of Sumatra. Journal of Tropical Ecology **16**, 253-270.

Ricklefs RE, Miller GL. 2000, Ecology. W.H. Freeman & Company, New York, USA.

Rodgers WA, Panwar HS. 1988. Planning a Protected Area Network in India. Vol I & II. The Report Wildlife Institute of India, Dehradun, India.

Saini DC. 2005. Flora of Bahraich District, Uttar Pradesh I - IV. Journal of Economic and Taxonomic Botany 29(3), 528-636.

Saini DC. 2005. Flora of Bahraich District, Uttar Pradesh - V. Journal of Economic and Taxonomic Botany **29(4)**, 843-920.

Shimada A, Wilson MW. 1985. Biological determinants of species diversity. Journal of Biogeography 12, 1-20.

Shrestha KK, Jha PK. 1997. Plant diversity and evaluation of conservation measures in the Royal Bardia National Park (RBNP). A report submitted to World Wildlife Fund Nepal Program. Kathmandu, Nepal.

Simpson EH. 1949. Measurement of diversity. Nature 163, 688.

Singh A, Reddy VS, Singh JS. 1995. Analysis of woody vegetation of Corbett National Park, India. Vegetation 120, 69-79.

Singh JS, Singh SP, Saxena AK, Ravat YS. 1984. The forest vegetation of Silent Valley. In Tropical Rain Forests - The Leeds Symposium, 25-52.

Singh L, Sharma B, Agrawal R, Puri S. 2005. Diversity and Dominance of a Tropical Moist Deciduous Forest in Achanakmar Wildlife Sanctuary. Bulletin of the National Institute of Ecology 15, 1-9.

Singh L, Singh JS. 1991. Species structure, dry matter dynamics and carbon flux of a dry tropical forest in India. Annals of Botany 68, 263-273.

Strasberg D. 1996. Diversity, size composition and spatial aggregation among trees on a 1 ha rain forest plot at La Reunion. Biodiversity and Conservation 5, 825-840.

Sundarapandian SM, Swamy PS. 1997. Plant biodiversity at low-elevation evergreen and moist deciduous forests at Kodayar (Western Ghats, India). International Journal of Ecology and Environmental Sciences 23, 363-379.

Sundarapandian SM, Swamy PS. 2000. Forest ecosystem structure and composition along an altitudinal gradient in the Western Ghats, South India. Journal of Tropical Forest Science 12, 104-123.

Supriya DL, Yadava PS. 2006. Floristic diversity assessment and vegetation analysis of tropical semievergreen forest of Manipur, north east India. Tropical Ecology 47(1), 89-98.

Swan JrFR. 1988. Tree distribution patterns in the Bukittimah nature reserve, Singapore. Garden Bulleten of Singapore 41, 59-81.

ter Braak CJF, Prentice IC. 1988. A theory of gradient analysis. Advances in Ecological Research 18, 271-313.

Tripathi KP. 2001. Ecology of a Rehabilitated Forest on Sodic Wasteland. Ph. D. Thesis, Lucknow University, Lucknow, India.

Tripathi KP, Singh B. 2009. Species diversity and vegetation structure across various strata in natural and plantation forests in Katerniaghat Wildlife Sanctuary, North India. Tropical Ecology 50(1), 191-200.

Tripathi KP, Tripathi S, Selven T, Kumar K, Singh KK, Mehrotra S, Pushpangadan P. 2004. Community structure and species diversity of saddle peak forests in Andaman Island. Tropical Ecology 45, 1-10.

Valencia R, Balslev H, Mino GCPY. 1994. High alpha-diversity in Amazonian Biodiversity and Conservation 3, 21-28.

Varghese AO, Menon ARR. 1998. Vegetation characteristics of southern secondary moist mixed deciduous forests of Agasthyamalai region of Kerala. Indian Journal of Forestry 21(4), 639-644.

Varghese AO, Menon ARR. 1999. Ecological niches and amplitudes of rare, threatened and endemic trees of Peppara Wildlife Sanctuary. Current Science 76, 1204-1208.

Venkateshwaran R, Parthasarathy N. 2005. Tree dry evergreen forests on the Coromandel Coast of India: structure, composition and human disturbance. Ecotropica 9, 45-58.

Visalakshi N. 1995. Vegetation analysis of two Tropical Dry Evergreen Forests in Southern India. Tropical Ecology **36**, 117-127.